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Article

# Glyphosate exposure and urinary oxidative stress biomarkers in the Agricultural Health Study

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#### **Abstract**

**Background:** Glyphosate is the most widely applied herbicide worldwide, and its use has been associated with increased risks of certain hematopoietic cancers in epidemiologic studies. Animal and in vitro experiments suggest that glyphosate may induce oxidative stress, a key characteristic of carcinogens; however, evidence in human populations remains scarce. We investigated associations between glyphosate exposure and urinary oxidative stress biomarkers in the Biomarkers of Exposure and Effect in Agriculture study, a molecular epidemiologic subcohort in the Agricultural Health Study.

**Methods:** This analysis included 268 male farmers selected based on self-reported recent and lifetime occupational glyphosate use and 100 age- and geography-matched male nonfarmers. Concentrations of glyphosate and oxidative stress biomarkers (8-hydroxy-2'-deoxyguanosine [8-OHdG], 8-iso-prostaglandin-F2 $\alpha$ , and malondialdehyde [MDA]) were quantified in first-morning-void urine. We performed multivariable linear regression to evaluate associations of urinary glyphosate and self-reported glyphosate use with each oxidative stress biomarker.

**Results:** Urinary glyphosate concentrations were positively associated with levels of 8-OHdG (highest vs lowest glyphosate quartile; geometric mean ratio = 1.15, 95% confidence interval = 1.03 to 1.28;  $P_{trend}$  = .02) and MDA (geometric mean ratio = 1.20, 95% confidence interval = 1.03 to 1.40;  $P_{trend}$  = .06) overall. Among farmers reporting recent glyphosate use (last 7 days), use in the previous day was also associated with statistically significantly increased 8-OHdG and MDA levels. Compared with nonfarmers, we observed elevated 8-iso-prostaglandin-F2 $\alpha$  levels among farmers with recent, high past 12-month, or high lifetime glyphosate use.

**Conclusions:** Our findings contribute to the weight of evidence supporting an association between glyphosate exposure and oxidative stress in humans and may inform evaluations of the carcinogenic potential of this herbicide.

Glyphosate is a broad-spectrum herbicide and crop desiccant. Since its commercialization in 1974, glyphosate has become the most widely applied agricultural pesticide in the United States and worldwide (1,2). As of 2012, glyphosate also ranked as the second most commonly used pesticide in US homes and gardens (2). Recently, nationally representative data from the National Health and Nutrition Examination Survey (2013-2014) suggest that approximately 80% of the general US population aged 6 years and older have detectable concentrations of glyphosate in their urine (3,4). Other biomonitoring studies suggest increasing exposure in the general population (5-7) and higher exposure among certain occupations, including farmers (8,9), with dermal contact being the major route of occupational exposure (10).

In 2015, the International Agency for Research on Cancer classified glyphosate as a probable human carcinogen (group 2A), citing limited epidemiologic evidence of an association with

non-Hodgkin lymphoma, sufficient evidence of carcinogenicity in experimental animals, and strong mechanistic evidence (mostly in animals and human cells) of genotoxicity and oxidative stress (11,12). However, the relationship between glyphosate exposure and risk of cancer, particularly lymphohematopoietic malignancies, remains inconclusive and controversial (13-15). The Agricultural Health Study (AHS), a prospective cohort of pesticide applicators in Iowa and North Carolina, recently reported a suggestive association between high lifetime use of glyphosate and increased risk of acute myeloid leukemia but not non-Hodgkin lymphoma or other cancers (16). Investigations of intermediate biomarkers of effect can provide timely evidence regarding the carcinogenic potential of this widely used herbicide (17).

Oxidative stress occurs when the production of reactive oxygen species (ROS) and other free radicals exceeds the body's antioxidant defense mechanisms, causing damage to DNA, proteins,

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and lipids (18). Although ROS form as part of normal cellular processes, they may also arise from exposure to exogenous agents, such as pesticides (18). The International Agency for Research on Cancer identified oxidative stress as a key characteristic of carcinogens (19-21), and accumulating evidence supports the role of oxidative stress in the pathogenesis of hematologic cancers (22,23). As such, assessment of glyphosate exposure in relation to markers of oxidative damage may provide insights into potential mechanisms underlying previously observed associations. Glyphosate has been shown to induce oxidative stress in human cells and animal models [reviewed previously (12,24)]; however, research in human populations is scarce. To our knowledge, only 4 studies among agricultural workers have investigated glyphosate exposure in relation to oxidative stress biomarkers (25-28), 2 of which reported positive associations (27,28). Notably, most of these studies had relatively small sample sizes; relied only on self-reported exposures; and lacked details on the timing, frequency, or history of glyphosate use.

In this investigation, we evaluated associations of urinary glyphosate concentrations and self-reported occupational glyphosate use with urinary biomarkers of oxidative DNA damage (8-hydroxy-2'-deoxyguanosine [8-OHdG]) and lipid peroxidation (8-iso-prostaglandin-F2α [8-isoprostane] and malondialdehyde [MDA]) among farmers and nonfarmers in the Biomarkers of Exposure and Effect in Agriculture (BEEA) study.

## **Methods**

# Study design and population

The BEEA study is a molecular epidemiologic subcohort nested within the AHS (29,30). Briefly, during 2010-2018, we enrolled 1681 male farmers from the AHS who were aged 50 years and older, resided in Iowa or North Carolina, had never been diagnosed with cancer (except nonmelanoma skin cancer), and completed questionnaires administered at AHS enrollment (1993-1997) and 2 follow-up interviews (1999-2003 and 2005-2010). We additionally enrolled 211 male nonfarmers from Iowa or North Carolina who were aged 50 years and older, had no history of cancer, and had not lived or worked on a farm or held a job that involved handling pesticides within the last 10 years or for more than 12 months since age 18 years. The nonfarmers were identified using voter registration lists and selected to have similar distributions as the BEEA farmers in terms of age, race and ethnicity (Black, White, or other [American Indian or Alaska Native, Asian, and Native Hawaiian or other Pacific Islander]), and state and county of residence (details described in Supplementary Methods, available online). At BEEA enrollment, study staff visited participants' homes to collect first-morning-void urine samples and administer a questionnaire soliciting information on demographics, lifestyle, and medical history, as well as use of specific pesticides (including recency and frequency of use) in the last 12 months. The BEEA protocol was approved by institutional review boards at the National Cancer Institute and other participating institutions. All participants provided written informed consent. The involvement of the Centers for Disease Control and Prevention laboratory did not constitute engagement in human subjects research.

For this investigation, we selected 4 subgroups of BEEA participants (N = 368) based on their reported glyphosate use: 1) recently exposed farmers with occupational glyphosate use during the 7 days prior to urine collection, regardless of lifetime use (n=98); 2) high lifetime-exposed farmers who were in the top 80th percentile of cumulative lifetime occupational glyphosate use but reported no use in the last 7 days (n = 70); 3) farming controls with minimal lifetime occupational glyphosate use (never used or have not used since after the 1999-2003 interview and in the lowest tertile of cumulative lifetime use) (n = 100); and 4) nonfarming controls with no home or garden use of glyphosate in the last 7 days (n = 100). The farming and nonfarming control groups were frequency matched to the glyphosate-exposed farmers (recently and high lifetime-exposed combined) by age (50-60, 61-70, older than 70 years), state (Iowa, North Carolina), and season of enrollment (April-September, October-March [off-season]). Details of the questionnaire-based glyphosate exposure assessment and study group definitions are described in Supplementary Methods (available online).

## Laboratory measurements

Urinary glyphosate concentrations were quantified at the Centers for Disease Control and Prevention (Atlanta, GA, USA) using ion chromatography isotope-dilution tandem mass spectrometry as described previously (31) and in Supplementary Methods (available online). The limit of detection (LOD) was  $0.2 \,\mu\text{g/L}$ ; concentrations below the LOD (n=45; 12.2%) were assigned a value of LOD/\2 (32). Oxidative stress biomarkers were quantified in urine using enzyme-linked immunosorbent (for 8-OHdG and 8-isoprostane) and thiobarbituric acid reactive substances (for MDA) assays at Cayman Chemical (Ann Arbor, MI, USA), as detailed in Supplementary Methods (available online). To account for urinary dilution, creatinine was quantified using an enzymatic method at the University of Minnesota Advanced Research and Diagnostic Laboratory (Minneapolis, MN, USA).

For each of these analyses, samples from participants in each of the 4 study groups were distributed evenly across all batches. To assess reproducibility of measurements, we included blinded quality control duplicate samples interspersed within and across batches. For glyphosate, 8-OHdG, 8-isoprostane, and MDA, respectively, the within-batch coefficients of variation were 2.9%, 16.8%, 15.8%, and 11.0%, and intraclass correlation coefficients were 0.997, 0.76, 0.80, and 0.95.

#### Statistical analysis

For our main analysis, we performed multivariable linear regression to evaluate associations between urinary glyphosate concentrations (quartiles) and natural log-transformed concentrations of each oxidative stress biomarker, overall and within each study group. Basic models adjusted for age (continuous) and urinary creatinine concentration (continuous; natural log transformed). Fully adjusted models additionally included study design-related variables (state, season, and time of urine collection), lifestyle and medical factors suggested to influence oxidative stress (body mass index, smoking status, alcohol consumption, recent nonsteroidal anti-inflammatory drug use, recent infection, history of diabetes, and history of hypertension and/or heart disease) (33-36), as well as occupational use of 2,4-dichlorophenoxyacetic acid (2,4-D), a commonly applied herbicide for which there is some prior evidence of associations with oxidative stress biomarkers (37) and the only pesticide used recently by more than 10% of farmers in this investigation. We reported associations as geometric mean ratios (GMRs) with 95% confidence intervals (CIs). GMRs were calculated by exponentiating the parameter estimates from linear regression models and interpreted as the ratio of geometric mean oxidative stress biomarker concentration for each glyphosate quartile relative to the lowest quartile. Tests for linear trend across quartiles were conducted by modeling within-quartile median values of glyphosate concentration as a

continuous variable. Additionally, we evaluated associations between continuous glyphosate concentration (log<sub>2</sub> transformed) and oxidative stress biomarkers.

We also evaluated associations between recent (last 7 days) occupational glyphosate use and oxidative stress biomarker concentrations compared with either farming or nonfarming controls as the referent category. Farmers with recent use were further classified by number of days since last use (≤1, 2-4, 5-7 days). To evaluate potential effects of longer-term or chronic exposure, we estimated associations of past 12-month and lifetime occupational glyphosate use (tertiles of intensity-weighted days of use; described in Supplementary Methods, available online) with oxidative stress biomarkers among glyphosateexposed farmers and compared with each of the 2 control groups.

We performed several sensitivity analyses to further assess potential confounding: restricting to White participants, Iowa residents, participants enrolled during farming season (April-September), or farmers without recent occupational 2,4-D use. To assess the impact of outliers and highly concentrated or diluted urine, we ran models excluding participants whose urinary oxidative stress biomarker concentrations were more than 3 standard deviations above the mean or those with creatinine concentrations outside the World Health Organization's reference range (30-300 mg/dL) (38). To assess the influence of nonoccupational exposure on findings, we also conducted analyses excluding participants with home and/or garden glyphosate use.

Statistical analyses were performed using SAS, v9.4 (Cary, NC, USA). All tests were 2-sided, with statistical significance evaluated at a P value less than .05.

## Results

Distributions of participant characteristics were generally similar across the 4 study groups, except for lower prevalence of diabetes and more common occupational 2,4-D use among recently and high lifetime glyphosate-exposed farmers (Table 1). Additionally, as expected, recently exposed farmers were more likely to be enrolled during farming season than other groups. We also noted lower prevalence of hypertension and/or heart disease and more common 2,4-D use among participants in higher urinary glyphosate quartiles (Supplementary Table 1, available online).

Urinary glyphosate concentrations were statistically significantly elevated among recently exposed farmers (geometric mean =  $0.89 \,\mu g/L$ ) compared with high lifetime-exposed farmers  $(0.59 \,\mu\text{g/L})$  and farming  $(0.46 \,\mu\text{g/L})$  and nonfarming  $(0.39 \,\mu\text{g/L})$  controls (all P < .01), whereas no statistically significant differences in 8-OHdG, 8-isoprostane, or MDA concentrations were observed across groups (Supplementary Table 2, available online). The 3 oxidative stress biomarkers were moderately correlated with one another (Spearman correlation coefficients =  $\sim$ 0.6-0.7), although correlations were attenuated for creatinine-corrected concentrations (Supplementary Table 3, available online).

Urinary concentrations of each oxidative stress biomarker increased with increasing quartiles of urinary glyphosate among all participants (Figure 1). In fully adjusted models, we observed statistically significant positive associations between urinary glyphosate and 8-OHdG (highest vs lowest quartile: GMR = 1.15 [95% CI = 1.03 to 1.28],  $P_{trend}\!=$  .02) and MDA (GMR = 1.20 [95% CI = 1.03 to 1.40],  $P_{trend} = .06$ ) but not 8-isoprostane (Table 2). Modest positive associations with 8-OHdG and MDA were also observed when urinary glyphosate was modeled as a continuous log<sub>2</sub>-transformed variable. Patterns of associations were generally similar when stratified by study group, particularly among recently exposed farmers (8-OHdG: GMR=1.23 [95% CI=0.97 to 1.57],  $P_{trend} = .03$ ; MDA: GMR = 1.19 [95% CI = 0.85 to 1.66], P<sub>trend</sub> = .43) and nonfarming controls (8-OHdG: GMR = 1.29 [95% CI = 1.00 to 1.67],  $P_{trend} = .06$ ; MDA: GMR = 1.17 [95% CI = 0.81 to 1.68], P<sub>trend</sub> = .39) (Supplementary Table 4, available online).

Table 3 presents fully adjusted associations between recent occupational glyphosate use (last 7 days) and urinary oxidative stress biomarkers (models only adjusted for age and creatinine shown in Supplementary Table 5, available online). Among recently exposed farmers, glyphosate use within 1 day (vs 5-7 days) of urine collection was associated with elevated concentrations of 8-OHdG (GMR = 1.20 [95% CI = 1.01 to 1.42]) and MDA (GMR = 1.28 [95% CI = 1.02 to 1.60]); comparisons with farming or nonfarming controls showed similar patterns but were not statistically significant. Furthermore, compared with nonfarmers, recent glyphosate use (regardless of further classification by days since last use) was associated with increased 8-isoprostane levels (GMR = 1.23 [95% CI = 1.03 to 1.47]).

In analyses examining longer-term occupational glyphosate use (Table 4 and Supplementary Table 6, available online), we found an association between high use in the last 12 months and elevated urinary 8-isoprostane levels in the fully adjusted model (tertile 3 of intensity-weighted days vs nonfarming controls: GMR = 1.21 [95% CI = 1.02 to 1.44]); a similar association was observed for high intensity-weighted lifetime days of use. We found no associations between these metrics and 8-OHdG or MDA. Stratified analyses showed similar results for recently exposed and high lifetime-exposed (but not recently) farmers (Supplementary Table 7, available online).

Each sensitivity analysis (described in "Methods") yielded similar results as our main analysis. Notably, urinary glyphosate remained positively associated with 8-OHdG and MDA across all analyses, and associations with MDA became slightly stronger after excluding participants with extreme MDA or creatinine values or restricting to those enrolled during farming season or those with no home or garden glyphosate use (Supplementary Table 8, available online).

#### Discussion

In this investigation among male farmers and demographically similar nonfarmers in Iowa and North Carolina, we observed associations between exposure to glyphosate and certain biomarkers of oxidative stress. Specifically, urinary glyphosate concentrations, as well as occupational glyphosate use in the previous day, were positively associated with urinary 8-OHdG and MDA. Compared with nonfarmers, we also observed elevated 8-isoprostane levels among farmers with occupational glyphosate use in the last 7 days and those with high past 12-month or lifetime use.

To our knowledge, only one previous study has evaluated occupational glyphosate exposure in relation to 8-OHdG, a promutagenic DNA lesion formed in response to ROS (25). Among 80 pesticide sprayers of an agricultural community in Greece, those who applied glyphosate at least once in the last spraying season were 1.5 times as likely to have high 8-OHdG levels (>75th percentile) in whole blood as those who did not; however, the association was based on univariate analysis and not statistically significant (25). Given that 8-OHdG reflects oxidative stressinduced DNA damage, our findings for 8-OHdG also support the genotoxic potential of glyphosate in humans (12) and strengthen existing evidence from studies that have reported associations

Table 1. Selected characteristics of glyphosate-exposed farmers and farming and nonfarming controls in the BEEA study

Characteristics <sup>a</sup>	Recently exposed (n = 98)	High lifetime-exposed (n = 70)	Farming controls (n = 100)	Nonfarming controls (n = 100)
Age, mean (SD), y	63.2 (8.3)	62.6 (9.4)	64.4 (10.1)	63.5 (8.7)
BMI, mean (SD), kg/m <sup>2</sup>	29.2 (5.3)	28.6 (4.4)	30.1 (5.9)	30.0 (6.1)
State	23.2 (3.3)	20.0 (4.4)	30.1 (3.3)	30.0 (0.1)
Iowa	77 (78.6)	49 (70.0)	76 (76.0)	76 (76.0)
North Carolina	,			
	21 (21.4)	21 (30.0)	24 (24.0)	24 (24.0)
Race	0 (0 0)	0 (0 0)	0 (0 0)	4 (4.0)
Black	0 (0.0)	0 (0.0)	2 (2.0)	1 (1.0)
White	97 (99.0)	70 (100.0)	96 (96.0)	98 (98.0)
Other <sup>b</sup>	1 (1.0)	0 (0.0)	2 (2.0)	1 (1.0)
Season of urine collection				
April-September	92 (93.9)	29 (41.4)	71 (71.0)	63 (63.0)
October-March (off-season)	6 (6.1)	41 (58.6)	29 (29.0)	37 (37.0)
Time of urine collection	, ,	, ,	•	,
Before 4:00 AM	22 (22.4)	8 (11.4)	14 (14.0)	16 (16.0)
4:00-5:59 AM	33 (33.7)	29 (41.4)	40 (40.0)	38 (38.0)
6:00 AM or later	43 (43.9)	33 (47.1)	46 (46.0)	46 (46.0)
Smoking status	15 (15.5)	33 (17.1)	10 (10.0)	10 (10.0)
Never	65 (66.3)	37 (52.9)	65 (65.0)	48 (48.0)
Former	28 (28.6)	31 (44.3)	32 (32.0)	45 (45.0)
Current				7 (7.0)
	5 (5.1)	2 (2.9)	3 (3.0)	7 (7.0)
Alcohol consumption (last 7 d) <sup>c</sup>	40 (40 0)	26 (51.4)	E4 (E4 O)	41 (41 0)
None	42 (42.9)	36 (51.4)	54 (54.0)	41 (41.0)
1-6 servings	41 (41.8)	20 (28.6)	27 (27.0)	31 (31.0)
≥7 servings	15 (15.3)	14 (20.0)	19 (19.0)	28 (28.0)
Recent NSAID use (last 7 d) <sup>d</sup>				
No	31 (31.6)	28 (40.0)	41 (41.0)	33 (33.0)
Yes	67 (68.4)	42 (60.0)	59 (59.0)	67 (67.0)
Recent infection (last 7 d) <sup>e</sup>				
No	90 (91.8)	60 (85.7)	88 (88.0)	88 (88.0)
Yes	8 (8.2)	10 (14.3)	12 (12.0)	12 (12.0)
History of diabetes	,	, ,	, ,	,
No	90 (91.8)	66 (94.3)	84 (84.0)	77 (77.0)
Yes	8 (8.2)	4 (5.7)	16 (16.0)	23 (23.0)
History of hypertension and/or	0 (0.2)	1 (3.7)	10 (10.0)	23 (23.3)
heart disease				
No	48 (49.0)	38 (54.3)	44 (44.0)	45 (45.0)
Yes			56 (56.0)	55 (55.0)
	50 (51.0)	32 (45.7)	36 (36.0)	33 (33.0)
Home and/or garden glypho-				
sate use	64 (60.0)	40 (50 0)	FO (FO O)	F.C. (F.C. O.)
Did not use in the last 12 mo	61 (62.2)	42 (60.0)	58 (58.0)	56 (56.0)
Used in the last 12 mo	37 (37.8)	28 (40.0)	42 (42.0)	44 (44.0)
Occupational 2,4-D use				
Did not use in the last 12 mo	22 (22.4)	22 (31.4)	87 (87.0)	_f
8-365 days ago	55 (56.1)	39 (55.7)	13 (13.0)	_f
≤7 days ago	21 (21.4)	9 (12.9)	0 (0.0)	_f

a Presented as frequencies and percentages (%) unless otherwise specified. 2,4-D=2,4-dichlorophenoxyacetic acid; BEEA = Biomarkers of Exposure and Effect in Agriculture; BMI = body mass index; NSAID = nonsteroidal anti-inflammatory drug.

American Indian or Alaska Native (n = 1) or not reported (n = 3).

Not applicable.

between glyphosate exposure and increased DNA damage, assessed as DNA strand breaks (39) or micronucleus formation (40). The associations we observed with MDA, an end product of ROS reaction with polyunsaturated fatty acids, further suggest that glyphosate may induce oxidative injury to cell membrane lipids (41) and are consistent with a recent study among 180 maize farmers in Thailand which reported a positive association between urinary glyphosate and serum MDA levels following glyphosate application (28). Conversely, another study of 93 Thai farmers who used glyphosate found no difference between preand postwork urinary MDA levels (26). Compared with most previous studies relying on self-reported use or geographic proximity to spraying, our analyses based on urinary glyphosate measurements may be more relevant to the effects of the internal glyphosate dose. Furthermore, although our study focused on occupational exposure in farmers, we also observed associations between urinary glyphosate and oxidative stress biomarkers, particularly 8-OHdG, among nonfarmers, suggesting these effects may apply more broadly to the general population who are primarily exposed through ingestion of contaminated food and water or residential applications (8). Two prior general population studies, one conducted among pregnant women (42) and the other among children (43), have examined glyphosate and oxidative stress biomarkers, both reporting positive associations for urinary glyphosate or its main metabolite, aminomethylphosphonic acid (AMPA).

Contrary to our results for 8-OHdG and MDA, we observed no associations between urinary glyphosate and 8-isoprostane,

Number of servings of alcoholic beverages in the last 7 days. One serving of an alcoholic beverage was defined as 12 fluid ounces of beer, 5 fluid ounces of wine, or 1.5 fluid ounces of hard liquor.

Use of any aspirin- or ibuprofen-containing products in the last 7 days. Having a cold, flu, or other infection during the last 7 days.

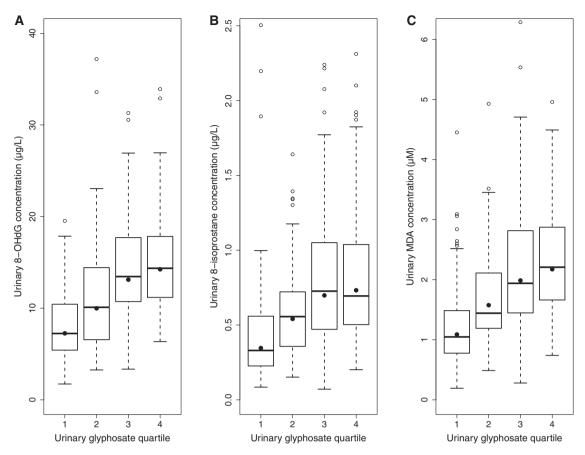


Figure 1. Box and whisker plots of urinary (A) 8-OHdG, (B) 8-isoprostane, and (C) MDA concentrations across quartiles of urinary glyphosate concentration among all 368 participants (quartile  $1 = \text{LOD-}0.289 \, \mu\text{g/L}$ ; quartile  $2 = 0.290 - 0.506 \, \mu\text{g/L}$ ; quartile  $3 = 0.507 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.507 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.934 - 0.933 \, \mu\text{g/L}$ ; quartile  $4 = 0.933 \, \mu\text{g/L$ 35.2 µg/L). The top and bottom edges of the boxes represent the upper (75th percentile) and lower (25th percentile) quartiles of the oxidative stress biomarker, respectively, and the whiskers above and below the boxes indicate the range of the data that lie within 1.5 times the IQR (ie, 1.5 IQR above the 75th percentile for the upper whisker and the minimum observed value for the lower whisker). The thick horizontal lines represent the median, and the solid circles represent the geometric mean concentration of the oxidative stress biomarker. For ease of visualization, the y-axis was truncated at a value of 3 IQRs above the 75th percentile for each oxidative stress biomarker. 8-isoprostane = 8-iso-prostaglandin-F2\alpha; 8-OHdG = 8-hydroxy-2'deoxyguanosine; IQR = interquartile range; LOD = limit of detection; MDA = malondialdehyde.

although levels of this marker were elevated among farmers with recent or high longer-term glyphosate use compared with nonfarmers. Like MDA, 8-isoprostane is a widely assessed biomarker of lipid peroxidation and has been suggested to be more stable within individuals over time than other oxidative stress markers (44,45). It is possible that the associations with 8-isoprostane reflect chronic effects of long-term glyphosate exposure, which may not be detected by markers reflecting immediate or shortterm responses to environmental stressors (46). This may also explain the associations we observed between urinary glyphosate, a marker of recent exposure given its short half-life in humans (approximately 5-10 hours) (47,48), as well as reported glyphosate use in the previous day, and 8-OHdG and MDA but not 8-isoprostane. Our findings are somewhat consistent with those from a study of 120 Brazilian agricultural and nonagricultural workers, where plasma 8-isoprostane levels were elevated among farmers reporting regular glyphosate use, although the specific time window of exposure was unclear (27). A study among 227 pregnant women in Puerto Rico reported suggestive positive associations between urinary glyphosate and 8-isoprostane (42); however, potential differences in glyphosate toxicokinetics may complicate comparisons across such different populations (49).

Findings from our study and several other human population studies to date (27,28,42,43) agree with in vitro and animal studies that have together provided strong evidence for the potential of glyphosate to induce oxidative stress (12,24). In particular, rodent studies have shown increased lipid peroxidation upon glyphosate exposure, as indicated by elevated MDA levels in blood or tissues of glyphosate-treated animals relative to untreated animals (12,24). In vitro and animal experiments evaluating glyphosate genotoxicity also suggest that glyphosate may contribute to formation of oxidative DNA adducts, including 8-OHdG (50). In addition, glyphosate has been shown to induce oxidative stress by altering antioxidant enzyme activity and/or levels of glutathione or other endogenous antioxidants in rodents (12). Using untargeted metabolomics profiling, a recent study in China identified statistically significant alterations in pathways related to glutathione metabolism among workers from glyphosate manufacturing facilities compared with participants from the general population, further suggesting that glyphosate exposure may disrupt the oxidant-antioxidant balance in humans

Oxidative stress has been implicated in lymphomagenesis and leukemogenesis (22,23), with in vivo evidence of oxidative stress-induced bone marrow injury upon exposure to known

**Table 2.** Associations between urinary glyphosate and oxidative stress biomarker concentrations in the BEEA study (n = 368)

Urinary glyphosate concentration (μg/L) No.	8-OHdG		8-isoprostane		MDA		
	No.	Age and creatinine adjusted <sup>a</sup> GMR <sup>c</sup> (95% CI)	Fully adjusted <sup>b</sup> GMR <sup>c</sup> (95% CI)	Age and creatinine adjusted <sup>a</sup> GMR <sup>c</sup> (95% CI)	Fully adjusted <sup>b</sup> GMR <sup>c</sup> (95% CI)	Age and creatinine adjusted <sup>a</sup> GMR <sup>c</sup> (95% CI)	Fully adjusted <sup>b</sup> GMR <sup>c</sup> (95% CI)
Quartile 1 ( <lod-0.289)< td=""><td>93</td><td>1 (Referent)</td><td>1 (Referent)</td><td>1 (Referent)</td><td>1 (Referent)</td><td>1 (Referent)</td><td>1 (Referent)</td></lod-0.289)<>	93	1 (Referent)	1 (Referent)	1 (Referent)	1 (Referent)	1 (Referent)	1 (Referent)
Quartile 2 (0.290-0.506)	91	1.02 (0.93 to 1.12)	1.04 (0.95 to 1.14)	1.03 (0.90 to 1.18)	1.01 (0.89 to 1.16)	1.08 (0.95 to 1.23)	1.11 (0.97 to 1.26)
Quartile 3 (0.507-0.933)	92	1.13 (1.02 to 1.24)	1.14 (1.03 to 1.26)	1.04 (0.90 to 1.19)	1.08 (0.94 to 1.25)	1.16 (1.01 to 1.33)	1.19 (1.03 to 1.37)
Quartile 4 (0.934-35.2)	92	1.12 (1.01 to 1.24)	1.15 (1.03 to 1.28)	0.96 (0.83 to 1.11)	1.02 (0.88 to 1.19)	1.16 (1.01 to 1.34)	1.20 (1.03 to 1.40)
P <sub>trend</sub> <sup>d</sup>		.03	.02	.33	` .89	.10	` .06
Continuous <sup>e</sup>	368	1.02 (0.99 to 1.05)	1.03 (1.00 to 1.06)	0.98 (0.94 to 1.02)	1.00 (0.96 to 1.04)	1.02 (0.98 to 1.06)	1.03 (0.99 to 1.07)

<sup>&</sup>lt;sup>a</sup> Adjusted for age (continuous; years) and natural log-transformed urinary creatinine concentration (continuous; mg/dL). 8-isoprostane = 8-iso-prostaglandin-F2 $\alpha$ ; 8-OHdG = 8-hydroxy-2'-deoxyguanosine; BEEA = Biomarkers of Exposure and Effect in Agriculture; CI = confidence interval; GMR = geometric mean ratio; LOD = limit of detection; MDA = malondialdehyde.

b Adjusted for age (continuous; years), natural log-transformed urinary creatinine concentration (continuous; mg/dL), state (lowa, North Carolina), season of urine collection (April-September, October-March), time of urine collection (before 4:00 AM, 4:00-5:59 AM, 6:00 AM or later), body mass index (continuous; kg/m²), smoking status (never, former, current), alcohol consumption (0, 1-6, ≥7 servings in the last 7 days), nonsteroidal anti-inflammatory drug use in the last 7 days (no, yes), infection in the last 7 days (no, yes), infection in the last 7 days (no, yes), infection in the last 7 days (no, yes), and occupational 2,4-dichlorophenoxyacetic acid use

<sup>(</sup>did not use in the last 12 months, 8-365 days ago, <7 days ago).

GMR represents the ratio of geometric mean urinary oxidative stress marker concentration compared with the reference group and was calculated by exponentiating the parameter estimate (e<sup>β</sup>) from linear regression model with natural log-transformed urinary oxidative stress marker concentration as the dependent variable.

d Calculated by modeling within-quartile median values as a continuous variable (0.204, 0.385, 0.677, and 1.435 µg/L for quartiles 1, 2, 3, and 4, respectively).

Per 1-unit increase in log2-transformed urinary glyphosate concentration, corresponding to a doubling in urinary glyphosate concentration.

Table 3. Associations between recent occupational glyphosate use (last 7 days) and urinary oxidative stress biomarker concentrations in the BEEA study

			Fully adjusted GMR (95% CI) <sup>a</sup>		
Glyphosate use	No.	Geometric mean concentration (95% CI)	Compared with nonfarming controls	Compared with farming controls	Among recently exposed only
8-OHdG (µg/L)					
Nonfarming controls	100	10.2 (9.2 to 11.3)	1 (Referent)	_	_
Farming controls	100	10.9 (9.8 to 12.2)		1 (Referent)	_
Recently exposed	98	10.7 (9.7 to 11.8)	0.98 (0.84 to 1.14)	0.97 (0.85 to 1.10)	_
Days since last use			· · · · ·	· ·	
5-7	42	9.4 (7.8 to 11.2)	0.92 (0.79 to 1.09)	0.92 (0.80 to 1.06)	1 (Referent)
2-4	35	11.0 (9.5 to 12.7)	0.98 (0.82 to 1.18)	0.96 (0.83 to 1.13)	1.04 (0.90 to 1.20)
≤1	21	13.5 (11.8 to 15.4)	1.11 (0.92 to 1.35)	1.08 (0.91 to 1.29)	1.20 (1.01 to 1.42)
8-isoprostane (μg/L)					
Nonfarming controls	100	0.53 (0.46 to 0.62)	1 (Referent)	_	_
Farming controls	100	0.55 (0.47 to 0.64)		1 (Referent)	_
Recently exposed	98	0.57 (0.50 to 0.64)	1.23 (1.03 to 1.47)	1.15 (0.96 to 1.38)	_
Days since last use		·	· · · · ·		
5-7	42	0.53 (0.43 to 0.64)	1.25 (1.03 to 1.51)	1.17 (0.95 to 1.44)	1 (Referent)
2-4	35	0.57 (0.46 to 0.70)	1.21 (0.98 to 1.49)	1.14 (0.91 to 1.43)	0.95 (0.80 to 1.12)
≤1	21	0.65 (0.50 to 0.85)	1.22 (0.96 to 1.53)	1.11 (0.86 to 1.43)	0.98 (0.81 to 1.19)
MDA (µM)		,	,	,	· · · · · · · · · · · · · · · · · · ·
Nonfarming controls	100	1.57 (1.38 to 1.78)	1 (Referent)	_	_
Farming controls	100	1.66 (1.47 to 1.88)	_	1 (Referent)	_
Recently exposed	98	1.69 (1.52 to 1.88)	0.98 (0.80 to 1.21)	1.03 (0.86 to 1.22)	_
Days since last use		,	,	,	
5-7	42	1.51 (1.26 to 1.80)	0.95 (0.76 to 1.18)	0.98 (0.80 to 1.20)	1 (Referent)
2-4	35	1.61 (1.37 to 1.90)	0.92 (0.72 to 1.17)	0.95 (0.77 to 1.19)	0.96 (0.79 to 1.17)
≤1	21	2.29 (1.83 to 2.88)	1.18 (0.91 to 1.55)	1.25 (0.98 to 1.60)	1.28 (1.02 to 1.60)

Adjusted for age (continuous; years), natural log-transformed urinary creatinine concentration (continuous; mg/dL), state (Iowa, North Carolina), season of urine collection (April-September, October-March), time of urine collection (before 4:00 AM, 4:00-5:59 AM, 6:00 AM or later), body mass index (continuous; kg/m²) smoking status (never, former, current), alcohol consumption (0, 1-6,  $\geq$ 7 servings in the last 7 days), nonsteroidal anti-inflammatory drug use in the last 7 days (no, yes), infection in the last 7 days (no, yes), history of diabetes (no, yes), history of hypertension and/or heart disease (no, yes), and occupational 2,4dichlorophenoxyacetic acid use (did not use in the last 12 months, 8-365 days ago, <7 days ago). 8-isoprostane = 8-iso-prostaglandin-F2a; 8-OHdG = 8-hydroxy-2'deoxyguanosine; BEEA = Biomarkers of Exposure and Effect in Agriculture; CI = confidence interval; GMR = geometric mean ratio; MDA = malondialdehyde.

carcinogens (52-54). Additionally, several case-control studies have reported higher urinary or blood levels of oxidative stress biomarkers, including 8-OHdG (55,56), MDA (57-59), and 8-isoprostane (60), among newly diagnosed hematologic cancer patients compared with healthy controls. As such, our findings provide mechanistic insights and biological plausibility for the potential role of glyphosate in the development of certain hematologic malignancies (12,16,61,62).

A distinctive feature of our study was the comprehensive exposure assessment, including urinary glyphosate measurements and well-characterized recent and lifetime occupational pesticide exposure histories, as well as the inclusion of farming and nonfarming controls. Other strengths included the larger sample size compared with other human studies, availability of information on a range of potential confounders, and detailed sensitivity analyses that confirmed the robustness of our findings. Our study also had several limitations. Given the crosssectional study design, biomarker measurements were from a single time point, which precluded the assessment of longitudinal associations, although we were able to explore potential temporal relationships based on timing and recency of glyphosate use. Although self-reported pesticide use may be subject to misclassification of exposure, reported use among AHS participants has been shown to be reliable (63,64), and questionnaire-assessed exposures and intensity metrics have been correlated with pesticide biomarkers in the AHS (65). Furthermore, we did not measure urinary concentrations of AMPA as an additional marker of exposure. However, AMPA is generally detected less frequently and at similar or lower concentrations than glyphosate because

of limited glyphosate metabolism in humans (8,49,66), and AMPA may form as a breakdown product of other phosphonatecontaining compounds (eg, detergents) besides glyphosate (67). Recent research also suggests that people may be primarily exposed to AMPA through food and water and to a lesser extent from metabolism of glyphosate (49). Nevertheless, future studies may consider assessing AMPA given experimental evidence of its potential role in oxidative stress (12) and associations observed with oxidative stress biomarkers in 2 recent general population studies (42,43). Lastly, although we measured 3 established and representative biomarkers of oxidative DNA damage or lipid peroxidation (21), they may not reflect the full extent of oxidative stress responses; future work using untargeted approaches may uncover additional oxidative stress-related metabolites or pathways associated with glyphosate exposure (51,68).

In conclusion, our findings suggest that glyphosate exposure may be positively associated with certain urinary biomarkers of oxidative stress. Although the observed associations mainly appear to reflect effects of recent occupational exposure, there was also some evidence of associations with longer-term exposure. Our study contributes to accumulating evidence supporting the role of glyphosate in oxidative stress among humans and provides insights into potential mechanisms underlying previously observed associations with some hematopoietic cancers. Future investigations including additional biomarkers of oxidative stress or other intermediate endpoints related to cancer development (eg, genotoxicity, epigenetic alterations) may further inform the evaluation of the carcinogenic potential of this herbicide.

Table 4. Associations of occupational glyphosate use in the last 12 months and cumulative lifetime occupational glyphosate use with urinary oxidative stress biomarker concentrations in the BEEA study

			Fully adjusted GMR (95% CI) <sup>a</sup>			
Glyphosate use	No.	Geometric mean concentration (95% CI)	Compared with nonfarming controls	Compared with farming controls	Glyphosate-exposed farmers only	
<del></del>						
8-OHdG (µg/L)	100	10.2 (0.2 to 11.2)	1 (Deferent)			
Nonfarming controls	100	10.2 (9.2 to 11.3) 10.9 (9.8 to 12.2)	1 (Referent)	1 (Referent)	_	
Farming controls Last 12-month use <sup>b</sup>	100	10.9 (9.8 to 12.2)	_	I (Referent)	_	
Tertile 1	56	11.0 (9.5 to 12.6)	1.04 (0.92 to 1.18)	1.02 (0.91 to 1.15)	1 (Referent)	
Tertile 1 Tertile 2	56	11.0 (9.5 to 12.6) 11.0 (9.7 to 12.5)	0.97 (0.85 to 1.11)	0.96 (0.85 to 1.08)	0.92 (0.82 to 1.04)	
Tertile 2 Tertile 3	56	11.0 (9.7 to 12.3) 11.2 (9.9 to 12.7)	1.06 (0.93 to 1.21)			
	30	11.2 (9.9 to 12.7)	.36	1.05 (0.93 to 1.18)	1.00 (0.89 to 1.13) .56	
P <sub>trend</sub> c Lifetime use <sup>d</sup>			.50	.40	.30	
Tertile 1	56	10.2 (8.8 to 11.9)	1.01 (0.89 to 1.16)	0.99 (0.88 to 1.12)	1 (Referent)	
Tertile 2	56	12.3 (11.1 to 13.7)	1.06 (0.93 to 1.20)	1.04 (0.93 to 1.17)	1.05 (0.93 to 1.18)	
Tertile 2 Tertile 3	55	10.6 (9.3 to 12.1)	0.99 (0.86 to 1.13)	0.98 (0.87 to 1.12)	0.97 (0.86 to 1.10)	
P <sub>trend</sub> <sup>c</sup>	55	10.6 (9.3 to 12.1)	0.99 (0.86 to 1.13) .75	.85	0.97 (0.86 to 1.10) .44	
8-isoprostane (µg/L)			./3	.63	.44	
Nonfarming controls	100	0.53 (0.46 to 0.62)	1 (Referent)			
	100	0.55 (0.47 to 0.64)	I (Referency	1 (Referent)	_	
Farming controls Last 12-month use <sup>b</sup>	100	0.55 (0.47 to 0.64)	_	I (Referent)	_	
Tertile 1	56	0.54 (0.44 to 0.65)	1.13 (0.97 to 1.32)	1.10 (0.93 to 1.29)	1 (Referent)	
Tertile 2	56	0.60 (0.51 to 0.70)	1.14 (0.96 to 1.34)	1.10 (0.93 to 1.29) 1.11 (0.93 to 1.33)	1.00 (0.87 to 1.15)	
Tertile 2	56	0.59 (0.49 to 0.70)	1.14 (0.96 to 1.34) 1.21 (1.02 to 1.44)	1.17 (0.98 to 1.40)	1.03 (0.90 to 1.19)	
P <sub>trend</sub> <sup>c</sup>	30	0.39 (0.49 to 0.70)	.08	.13	.63	
Lifetime use <sup>d</sup>			.08	.13	.03	
Tertile 1	56	0.53 (0.45 to 0.64)	1.19 (1.01 to 1.41)	1.14 (0.95 to 1.35)	1 (Referent)	
Tertile 2	56	0.61 (0.52 to 0.72)	1.19 (1.01 to 1.41) 1.10 (0.94 to 1.29)	1.08 (0.91 to 1.28)	0.94 (0.82 to 1.10)	
Tertile 3	55	0.57 (0.48 to 0.69)	1.10 (0.94 to 1.29) 1.20 (1.00 to 1.42)	1.17 (0.97 to 1.40)	1.03 (0.88 to 1.20)	
P <sub>trend</sub> <sup>c</sup>	33	0.37 (0.48 to 0.69)	.19	.20	.51	
MDA (µM)			.19	.20	.51	
Nonfarming controls	100	1.57 (1.38 to 1.78)	1 (Referent)	_	_	
Farming controls	100	1.66 (1.47 to 1.88)	i (Referency	1 (Referent)		
Last 12-month use <sup>b</sup>	100	1.00 (1.47 to 1.88)		i (Referency		
Tertile 1	56	1.65 (1.44 to 1.88)	1.01 (0.85 to 1.20)	1.00 (0.85 to 1.18)	1 (Referent)	
Tertile 2	56	1.67 (1.44 to 1.95)	0.94 (0.78 to 1.13)	0.95 (0.80 to 1.13)	0.92 (0.78 to 1.07)	
Tertile 3	56	1.75 (1.52 to 2.01)	1.07 (0.89 to 1.29)	1.06 (0.90 to 1.26)	1.03 (0.88 to 1.20)	
P <sub>trend</sub> <sup>c</sup>	50	1.73 (1.32 to 2.01)	.26	.34	.40	
Lifetime use <sup>d</sup>			.20	.51	.40	
Tertile 1	56	1.62 (1.38 to 1.91)	1.01 (0.84 to 1.22)	1.02 (0.86 to 1.20)	1 (Referent)	
Tertile 2	56	1.72 (1.51 to 1.96)	0.98 (0.83 to 1.17)	0.98 (0.83 to 1.15)	0.95 (0.81 to 1.12)	
Tertile 3	55	1.72 (1.51 to 1.90) 1.74 (1.52 to 1.98)	1.05 (0.87 to 1.28)	1.04 (0.87 to 1.25)	1.04 (0.88 to 1.24)	
P <sub>trend</sub> <sup>c</sup>	رر	1.7 + (1.32 to 1.30)	.59	.67	.48	
¹ trend			دد.	.07	.40	

Adjusted for age (continuous; years), natural log-transformed urinary creatinine concentration (continuous; mg/dL), state (Iowa, North Carolina), season of urine collection (April-September, October-March), time of urine collection (before 4:00 AM, 4:00-5:59 AM, 6:00 AM or later), body mass index (continuous; kg/m²). smoking status (never, former, current), alcohol consumption (0, 1-6,  $\geq$ 7 servings in the last 7 days), nonsteroidal anti-inflammatory drug use in the last 7 days (no, yes), infection in the last 7 days (no, yes), history of diabetes (no, yes), history of hypertension and/or heart disease (no, yes), and occupational 2,4-dichlorophenoxyacetic acid use (did not use in the last 12 months, 8-365 days ago,  $\leq$ 7 days ago). 8-isoprostane = 8-iso-prostaglandin-F2 $\alpha$ ; 8-OHdG = 8-hydroxy-2'-deoxyguanosine; BEEA = Biomarkers of Exposure and Effect in Agriculture; CI = confidence interval; GMR = geometric mean ratio; MDA = malondialdehyde.

b Intensity-weighted days of occupational glyphosate use in the last 12 months among glyphosate-exposed farmers (tertile 1 = 0-512; tertile 2 = >512-1320; tertile 3 = >1320-11375), calculated by multiplying the number of days of use in the last 12 months by an exposure intensity score that accounts for factors known to influence pesticide exposure

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Author contributions: Vicky C. Chang (Formal analysis, Methodology; Writing - original draft); Gabriella Andreotti (Conceptualization; Writing - review & editing); Maria Ospina (Methodology; Writing – review & editing); Christine G. Parks (Writing - review & editing); Danping Liu (Formal analysis; Writing - review & editing); Joseph J. Shearer (Conceptualization;

Calculated by modeling within-tertile median values as continuous variables.

d Intensity-weighted lifetime days of occupational glyphosate use among glyphosate-exposed farmers (tertile 1 = 1321-11440; tertile 2 = >11440-23071; tertile 3 = >23071-244237), calculated by multiplying the total number of lifetime days of use by an exposure intensity score that accounts for factors known to influence pesticide exposure. One participant had missing data on lifetime glyphosate use and was excluded from the analysis.

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# Data availability

The data underlying this investigation will be provided upon request as described for the BEEA subcohort on the AHS website: https://aghealth.nih.gov/collaboration/studies.html. For more information on the process of requesting access to these data, including any administrative and human subjects requirements, please contact the Principal Investigator of the BEEA study (Dr Jonathan Hofmann) at hofmannjn@mail.nih.gov.

#### References

- 1. Benbrook CM. Trends in glyphosate herbicide use in the United States and globally. Environ Sci Eur. 2016;28(1):3.
- 2. Atwood D, Paisley-Jones C. Pesticides Industry Sales and Usage: 2008–2012 Market Estimates. Washington, DC: U.S. Environmental Protection Agency; 2017.
- Centers for Disease Control and Prevention. National Health and Nutrition Examination Survey. 2013-2014 data documentation, codebook, and frequencies: glyphosate (GLYP) - urine (SSGLYP\_H). 2022. https://wwwn.cdc.gov/Nchs/Nhanes/2013-2014/SSGLYP\_H.htm. Accessed July 29, 2022.
- 4. Centers for Disease Control and Prevention. National report on human exposure to environmental chemicals: what's new? urinary glyphosate (N-(Phosphonomethyl)glycine) data tables. 2022. https://www.cdc.gov/exposurereport/whats\_new\_071922 \_1.html. Accessed July 29, 2022.
- 5. Conrad A, Schröter-Kermani C, Hoppe HW, et al. Glyphosate in German adults - time trend (2001 to 2015) of human exposure to a widely used herbicide. Int J Hyg Environ Health. 2017;220(1): 8-16.

- 6. Mills PJ, Kania-Korwel I, Fagan J, et al. Excretion of the herbicide glyphosate in older adults between 1993 and 2016. JAMA. 2017;318(16):1610-1611.
- 7. Nomura H, Hamada R, Wada K, et al. Temporal trend and crosssectional characterization of urinary concentrations of glyphosate in Japanese children from 2006 to 2015. Int J Hyg Environ Health. 2022;242:113963.
- Gillezeau C, van Gerwen M, Shaffer RM, et al. The evidence of human exposure to glyphosate: a review. Environ Health. 2019;18(1):2.
- Grau D, Grau N, Gascuel Q, et al. Quantifiable urine glyphosate levels detected in 99% of the French population, with higher values in men, in younger people, and in farmers. Environ Sci Pollut Res Int. 2022;29(22):32882-32893.
- 10. Connolly A, Coggins MA, Galea KS, et al. Evaluating glyphosate exposure routes and their contribution to total body burden: a study among amenity horticulturalists. Ann Work Expo Health. 2019;63(2):133-147.
- 11. Guyton KZ, Loomis D, Grosse Y, et al.; for the International Agency for Research on Cancer Monograph Working Group, IARC, Lyon, France. Carcinogenicity of tetrachlorvinphos, parathion, malathion, diazinon, and glyphosate. Lancet Oncol. 2015;16(5):490-491.
- 12. International Agency for Research on Cancer. Some organophosphate insecticides and herbicides - Glyphosate. In: IARC Monographs on the Evaluation of Carcinogenic Risks to Humans, Volume 112. Lyon, France: International Agency for Research on Cancer; 2015:321-412.
- 13. Chang ET, Delzell E. Systematic review and meta-analysis of glyphosate exposure and risk of lymphohematopoietic cancers. J Environ Sci Health B. 2016;51(6):402-434.
- 14. Zhang L, Rana I, Shaffer RM, et al. Exposure to glyphosate-based herbicides and risk for non-Hodgkin lymphoma: a metaanalysis and supporting evidence. Mutat Res Rev Mutat Res. 2019:781:186-206.
- 15. Kogevinas M. Probable carcinogenicity of glyphosate. BMJ. 2019;365:11613.
- 16. Andreotti G, Koutros S, Hofmann JN, et al. Glyphosate use and cancer incidence in the Agricultural Health Study. J Natl Cancer Inst. 2018;110(5):509-516.
- 17. Ward EM. Glyphosate use and cancer incidence in the Agricultural Health Study: an epidemiologic perspective. J Natl Cancer Inst. 2018;110(5):446-447.
- 18. Rahal A, Kumar A, Singh V, et al. Oxidative stress, prooxidants, antioxidants: the interplay. Biomed 2014;2014:761264.
- 19. Smith MT, Guyton KZ, Gibbons CF, et al. Key characteristics of carcinogens as a basis for organizing data on mechanisms of carcinogenesis. Environ Health Perspect. 2016;124(6):713-721.
- 20. Guyton KZ, Rusyn I, Chiu WA, et al. Application of the key characteristics of carcinogens in cancer hazard identification. Carcinogenesis. 2018;39(4):614-622.
- 21. Smith MT, Guyton KZ, Kleinstreuer N, et al. The key characteristics of carcinogens: relationship to the hallmarks of cancer, relevant biomarkers, and assays to measure them. Cancer Epidemiol Biomarkers Prev. 2020;29(10):1887-1903.
- 22. Sillar JR, Germon ZP, DeIuliis GN, Dun MD. The role of reactive oxygen species in acute myeloid leukaemia. Int J Mol Sci. 2019;20(23):6003.
- 23. Rodríguez-García A, García-Vicente R, Morales ML, et al. Protein carbonylation and lipid peroxidation in hematological malignancies. Antioxidants (Basel). 2020;9(12):1212.

- 24. Wang X, Lu Q, Guo J, et al. Oxidative stress and metabolism: a mechanistic insight for glyphosate toxicology. Annu Rev Pharmacol Toxicol. 2022;62(1):617-639.
- 25. Koureas M, Tsezou A, Tsakalof A, et al. Increased levels of oxidative DNA damage in pesticide sprayers in Thessaly Region (Greece). Implications of pesticide exposure. Sci Total Environ. 2014;496:358-364.
- 26. Intayoung U, Wunnapuk K, Kohsuwan K, et al. Effect of occupational exposure to herbicides on oxidative stress in sprayers. Saf Health Work. 2021;12(1):127-132.
- 27. de Souza Espindola Santos A, Parks CG, Senna MM, et al. Exposure to pesticides and oxidative stress in Brazilian agricultural communities. Biomarkers. 2021;26(6):539-547.
- 28. Sidthilaw S, Sapbamrer R, Pothirat C, et al. Effects of exposure to glyphosate on oxidative stress, inflammation, and lung function in maize farmers, Northern Thailand. BMC Public Health. 2022;22(1):1343.
- 29. Alavanja MC, Sandler DP, McMaster SB, et al. The Agricultural Health Study. Environ Health Perspect. 1996;104(4):362-369.
- 30. Hofmann JN, Beane Freeman LE, Lynch CF, et al. The Biomarkers of Exposure and Effect in Agriculture (BEEA) study: rationale, design, methods, and participant characteristics. J Toxicol Environ Health A. 2015;78(21-22):1338-1347.
- 31. Schütze A, Morales-Agudelo P, Vidal M, et al. Quantification of glyphosate and other organophosphorus compounds in human urine via ion chromatography isotope dilution tandem mass spectrometry. Chemosphere. 2021;274:129427.
- 32. Hornung RW, Reed LD. Estimation of average concentration in the presence of nondetectable values. Appl Occup Environ Hyg. 1990;5(1):46-51.
- 33. Block G, Dietrich M, Norkus EP, et al. Factors associated with oxidative stress in human populations. Am J Epidemiol. 2002;156(3):274-285.
- 34. Keaney JF Jr, Larson MG, Vasan RS, et al.; for the Framingham Study. Obesity and systemic oxidative stress: clinical correlates of oxidative stress in the Framingham Study. Arterioscler Thromb Vasc Biol. 2003;23(3):434-439.
- 35. Black CN, Bot M, Scheffer PG, et al. Sociodemographic and lifestyle determinants of plasma oxidative stress markers 8-OHdG and F2-isoprostanes and associations with metabolic syndrome. Oxid Med Cell Longev. 2016;2016:7530820.
- 36. Sharifi-Rad M, Anil Kumar NV, Zucca P, et al. Lifestyle, oxidative stress, and antioxidants: back and forth in the pathophysiology of chronic diseases. Front Physiol. 2020;11:694.
- 37. Lerro CC, Beane Freeman LE, Portengen L, et al. A longitudinal study of atrazine and 2,4-D exposure and oxidative stress markers among Iowa corn farmers. Environ Mol Mutagen. 2017;58(1):30-38.
- 38. World Health Organization. Biological Monitoring of Chemical Exposure in the Workplace: Guidelines. Volume 1. Geneva, Switzerland: World Health Organization; 1996.
- 39. Paz-y-Miño C, Sánchez ME, Arévalo M, et al. Evaluation of DNA damage in an Ecuadorian population exposed to glyphosate. Genet Mol Biol. 2007;30(2):456-460.
- 40. Bolognesi C, Carrasquilla G, Volpi S, et al. Biomonitoring of genotoxic risk in agricultural workers from five Colombian regions: association to occupational exposure to glyphosate. J Toxicol Environ Health A. 2009;72(15-16):986-997.
- 41. Ayala A, Muñoz MF, Argüelles S. Lipid peroxidation: production, metabolism, and signaling mechanisms of malondialdehyde 4-hydroxy-2-nonenal. Oxid Med Cell 2014;2014:360438.

- 42. Eaton JL, Cathey AL, Fernandez JA, et al. The association between urinary glyphosate and aminomethyl phosphonic acid with biomarkers of oxidative stress among pregnant women in the PROTECT birth cohort study. Ecotoxicol Environ Saf. 2022:233:113300.
- 43. Makris KC, Efthymiou N, Konstantinou C, et al. Oxidative stress of glyphosate, AMPA and metabolites of pyrethroids and chlorpyrifos pesticides among primary school children in Cyprus. Environ Res. 2022;212(pt B):113316.
- 44. Il'yasova D, Scarbrough P, Spasojevic I. Urinary biomarkers of oxidative status. Clin Chim Acta. 2012;413(19-20):1446-1453.
- 45. Wu X, Cai H, Xiang YB, et al. Intra-person variation of urinary biomarkers of oxidative stress and inflammation. Cancer Epidemiol Biomarkers Prev. 2010;19(4):947-952.
- 46. Monaghan P, Metcalfe NB, Torres R. Oxidative stress as a mediator of life history trade-offs: mechanisms, measurements and interpretation. Ecol Lett. 2009;12(1):75-92.
- 47. Connolly A, Jones K, Basinas I, et al. Exploring the half-life of glyphosate in human urine samples. Int J Hyg Environ Health. 2019;222(2):205-210.
- 48. Zoller O, Rhyn P, Zarn JA, et al. Urine glyphosate level as a quantitative biomarker of oral exposure. Int J Hyg Environ Health. 2020;228:113526.
- 49. Connolly A, Coggins MA, Koch HM. Human biomonitoring of glyphosate exposures: state-of-the-art and future research challenges. Toxics. 2020;8(3):60.
- 50. Bolognesi C, Bonatti S, Degan P, et al. Genotoxic activity of glyphosate and its technical formulation roundup. J Agric Food Chem. 1997;45(5):1957-1962.
- 51. Zhang Q, Liu X, Gao M, et al. The study of human serum metabolome on the health effects of glyphosate and early warning of potential damage. Chemosphere. 2022;298:134308.
- 52. McHale CM, Zhang L, Smith MT. Current understanding of the mechanism of benzene-induced leukemia in humans: implications for risk assessment. Carcinogenesis. 2012;33(2):240-252.
- 53. Zhang Y, Liu X, McHale C, et al. Bone marrow injury induced via oxidative stress in mice by inhalation exposure to formaldehyde. PLoS One. 2013;8(9):e74974.
- 54. Wei C, Wen H, Yuan L, et al. Formaldehyde induces toxicity in mouse bone marrow and hematopoietic stem/progenitor cells and enhances benzene-induced adverse effects. Arch Toxicol. 2017;91(2):921-933.
- 55. Honda M, Yamada Y, Tomonaga M, et al. Correlation of urinary 8-hydroxy-2'-deoxyguanosine (8-OHdG), a biomarker of oxidative DNA damage, and clinical features of hematological disorders: a pilot study. Leuk Res. 2000;24(6):461-468.
- 56. Zhou FL, Zhang WG, Wei YC, et al. Involvement of oxidative stress in the relapse of acute myeloid leukemia. J Biol Chem. 2010;285(20):15010-15015.
- 57. Rasool M, Farooq S, Malik A, et al. Assessment of circulating biochemical markers and antioxidative status in acute lymphoblastic leukemia (ALL) and acute myeloid leukemia (AML) patients. Saudi J Biol Sci. 2015;22(1):106-111.
- 58. Hlavackova A, Vydra J, Chrastinova L, et al. Alteration of serum malondialdehyde level as biomarker of oxidative stress during acute myeloid leukemia treatment. Blood. 2019;134(suppl 1):5181.
- 59. Tsamesidis I, Pantaleo A, Pekou A, et al. Correlation of oxidative stress biomarkers and hematological parameters in blood cancer patients from Sardinia, Italy. Int J Hematol Oncol Stem Cell Res. 2019;13(2):49-57.

- 60. Faridvand Y, Oskuyi AE, Khadem-Ansari MH. Serum 8-isoprostane levels and paraoxonase 1 activity in patients with stage I multiple myeloma. Redox Rep. 2016;21(5):204-208.
- 61. Pahwa M, Beane Freeman LE, Spinelli JJ, et al. Glyphosate use and associations with non-Hodgkin lymphoma major histological sub-types: findings from the North American Pooled Project. Scand J Work Environ Health. 2019;45(6):600-609.
- 62. Leon ME, Schinasi LH, Lebailly P, et al. Pesticide use and risk of non-Hodgkin lymphoid malignancies in agricultural cohorts from France, Norway and the USA: a pooled analysis from the AGRICOH consortium. Int J Epidemiol. 2019;48(5):1519-1535.
- 63. Blair A, Tarone R, Sandler D, et al. Reliability of reporting on lifestyle and agricultural factors by a sample of participants in the Agricultural Health Study from Iowa. Epidemiology. 2002;13(1):94-99.
- 64. Hoppin JA, Yucel F, Dosemeci M, et al. Accuracy of self-reported pesticide use duration information from licensed pesticide

- applicators in the Agricultural Health Study. J Expo Anal Environ Epidemiol. 2002;12(5):313-318.
- 65. Thomas KW, Dosemeci M, Coble JB, et al. Assessment of a pesticide exposure intensity algorithm in the Agricultural Health Study. J Expo Sci Environ Epidemiol. 2010;20(6): 559-569.
- 66. Agency for Toxic Substances and Disease Registry (ATSDR). Toxicological Profile for Glyphosate. Atlanta, GA: U.S. Department of Health and Human Services, Public Health Service; 2020.
- 67. Grandcoin A, Piel S, Baurès E. AminoMethylPhosphonic acid (AMPA) in natural waters: its sources, behavior and environmental fate. Water Res. 2017;117:187-197.
- 68. Andrisic L, Dudzik D, Barbas C, et al. Short overview on metabolomics approach to study pathophysiology of oxidative stress in cancer. Redox Biol. 2018:14:47-58.